

# The changing ecology of primate parasites: insights from wild-captive comparisons

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1 2 3	1	The changing ecology of primate parasites: insights from wild-captive comparisons
4 5	2	Running headline: Changing ecology of primate parasites
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31 32 33	18	
34 35 36	19	Abstract
30 37 38	20	Host movements, including migrations or range expansions, are known to influence parasite
39 40	21	communities. Transitions to captivity—a rarely studied yet widespread human-driven host
41 42	22	movement-can also change parasite communities, in some cases leading to pathogen spillover
43 44 45	23	among wildlife species, or between wildlife and human hosts. We compared parasite species
46 47	24	richness between wild and captive populations of 22 primate species, including macro-
48 49	25	(helminths and arthropods) and micro-parasites (viruses, protozoa, bacteria, and fungi). We
50 51 52	26	predicted that captive primates would have only a subset of their native parasite community, and
52 53 54	27	would possess fewer parasites with complex life cycles requiring intermediate hosts or vectors.
55 56 57 58	28	We further predicted that captive primates would have parasites transmitted by close contact and 1

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29	environmentally-including those shared with humans and other animals such as commensals
30	and pests. We found that the composition of primate parasite communities shifted in captive
31	populations, especially due to turnover (parasites detected in captivity but not reported in the
32	wild), but with some evidence of nestedness (holdovers from the wild). Because of the high
33	degree of turnover, we found no significant difference in overall parasite richness between
34	captive and wild primates. Vector-borne parasites were less likely to be found in captivity,
35	whereas parasites transmitted through either close or non-close contact, including through fecal-
36	oral transmission, were more likely to be newly detected in captivity. These findings identify
37	parasites that require monitoring in captivity, and raise concerns about the introduction of novel
38	parasites to potentially susceptible wildlife populations during reintroduction programs.
39 40	Keywords: host-parasite interactions, nestedness, parasite species richness, turnover, zoonosis Research highlights:
41 42 43 44 45 46 47 48 49 50 51	<ul> <li>Changes in host environments—from wild to captive—can lead to changes in parasitism; studying these changes can inform captive wildlife management, wildlife relocation programs, and zoonotic disease risk assessment.</li> <li>Comparing 22 species of primates, we found high parasite species turnover in captive hosts, but no overall difference in parasite richness between wild and captive populations. Captive primates had fewer vector-borne parasites, and appeared to gain parasites transmitted via environmental exposure and close contact.</li> <li>Parasitism is an important consideration in translocating primates from the wild, and in reintroduction programs, owing to the potential for novel parasite transfers into human and wild primate populations.</li> </ul>

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52	1.	INTRODUCTION	

53 When moving into a new habitat, hosts can lose some parasite species, retain others, and acquire 54 new ones from novel environments or hosts. Transitions to new environments can occur through 55 multiple mechanisms, including dispersal and migration (e.g., Altizer, Bartel, & Han, 2011), the 56 unintentional anthropogenic introduction of plants and animals, and by intentional translocation 57 of wildlife by humans (Chomel, Belotto, & Meslin, 2007; Snyder et al., 1996; Wolfe et al., 1998). Capturing wild animals and moving them into captivity is a form of translocation that 58 occurs for a variety of reasons, including pet and wildlife trade, to acquire animals for captive 59 60 research, and for conservation purposes (Mittermeier, Konstant, & Mast, 1994; Smith et al., 2009). 61

An ecological understanding of how parasites of captive populations differ from their 62 63 wild counterparts is important for investigating fundamental questions in wildlife disease ecology, and also for evaluating health outcomes of captivity to inform captive breeding 64 programs and efforts to re-introduce captive individuals into the wild (Cunningham, 1996; 65 66 Hudson, Dobson, & Lafferty, 2006; Lyles & Dobson, 1993). Here, we use the ecological 67 definition of a parasite as any infectious agent that lives in or on a host, at some cost to that host, 68 including micro-parasites (viruses, bacteria, fungi and protozoa) and macro-parasites (helminths 69 and arthropods). The transition of hosts from the wild to captivity has important parallels with 70 parasite dynamics observed in migratory animals and in exotic species introductions, both of 71 which can reduce infection risk (Altizer et al., 2011; Torchin, Lafferty, Dobson, McKenzie, & Kuris, 2003; Torchin & Mitchell, 2004). Specifically, migratory animals can escape from 72 73 parasites in their breeding range as they move to their winter range, and heavily infected 74 individuals may die during strenuous migrations, lowering parasite prevalence (Altizer et al.,

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2011; Altizer, Hobson, Davis, De Roode, & Wassenaar, 2015). Similarly, when hosts are introduced into new environments, they often lose parasite species present in their native range and experience lower parasite burdens, which facilitates their invasion (Torchin et al. 2003; Mitchell and Power 2003).

79 Wildlife might lose parasites during three stages of transition from natural habitats to 80 captivity: (i) collection from the wild, because captured individuals likely harbor only a subset of parasites from the original wild population (Torchin et al., 2003), (ii) transport to captivity, 81 because the stress of transport and acute infections might cause some infected animals to die 82 83 (e.g., Kock, Mihok, Wambua, Mwanzia, & Saigawa, 1999; Lafferty & Holt, 2003; Scope, Filip, Gabler, & Resch, 2002), and (iii) establishment in captivity, where parasites from the native 84 range might be lost due to housing conditions that are not conducive to pathogen transmission. 85 86 Specialized parasites capable of infecting only one or a few host species might be more likely to be lost in captivity (Lyles & Dobson, 1993), whereas generalist parasites that can infect a broad 87 range of host species might tend to persist in captive environments that house multiple species. 88 89 Similarly, captive animals might disproportionately lose parasites with complex life cycles if 90 vectors or intermediate hosts necessary for transmission are rare or absent from captive settings 91 (Torchin et al., 2003). Finally, captive animals often receive medical treatment to reduce parasite loads, such as with anti-helminthic drugs, antibiotics, or vaccines (Munene et al., 1998), 92 potentially resulting in further declines in parasite diversity. 93

Alongside the loss of parasites from the native wild environment, captivity could
facilitate the acquisition of novel parasites. Stressful conditions in captive settings might
suppress host immunity, leaving captive hosts susceptible to new infections (Fowler, 1986; Lyles
& Dobson, 1993; Mason, 2010). Captive animals might also gain parasites when their housing

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98	facilitates close proximity to other host species not encountered in the wild, including
99	domesticated species and humans (Lyles & Dobson, 1993). This is particularly important if two
100	or more host species are phylogenetically similar, which has been shown to predict parasite
101	sharing in wild populations (Cooper, Griffin, Franz, Omotayo, & Nunn, 2012; Gilbert & Webb,
102	2007). Captive animals might also acquire parasites through exposure to new intermediate hosts
103	or vectors, especially when housed outdoors (Pung, Spratt, Clark, Norton, & Carter, 1998;
104	Ratterree et al., 2003). If captive animals are re-introduced, they have the potential to transmit
105	novel pathogens acquired in captivity to wild individuals (Hatcher, Dick, & Dunn, 2012; Lyles &
106	Dobson, 1993), posing risks to wild populations.
107	Primates are an especially important host group in which to consider parasite differences
108	between wild and captive environments. The risk of parasite spillover from captive nonhuman
109	primates to humans is substantial in zoos, laboratories, and rescue centers. In this context,
110	captive primates harbor many different parasites (Brack, 2012; Johnson-Delaney, 2009; Lyles &
111	Dobson, 1993; McPherson, 2013), some of which can infect humans and other animals (Ballou,
112	1993; Gyuranecz et al., 2009; Jones-Engel et al., 2004; McPherson, 2013; Weigler, 1992). For
113	example, research indicates that captive primates might be responsible for Leptospira and simian
114	foamy virus infections among zookeepers (Romero, Astudillo, Sánchez, González, & Varela,
115	2011; Sandstrom et al., 2000). Similarly, monkeys and an employee tested seropositive for
116	Reston Ebola virus at a quarantine facility in Virginia, and a young lab worker died tragically
117	after acquiring herpes B virus from a macaque at a primate research center (CDC, 1998; Miranda
118	et al., 1999). Thus, understanding the ecology of captive primate parasites is important to both
119	human and nonhuman animal health.

We compared parasite diversity between wild and captive populations using a new database of 22 captive primate species that have been sampled well for parasites in wild populations. To investigate population-level differences in exposure and susceptibility to parasites, we compared parasite species richness (PSR), or the total number of parasite species per host (Fréderic Bordes & Morand, 2009; Frédéric Bordes & Morand, 2011). Based on findings for invasive species (Mitchell & Power, 2003; Torchin et al., 2003; Torchin & Mitchell, 2004), we predicted that captive primates would have lower PSR than their wild counterparts. We investigated changes in the composition of parasite communities in wild and captive primates using beta diversity (Koleff, Gaston, & Lennon, 2003). We quantified both nestedness and turnover of parasite communities (Andrés Baselga, 2010), where nestedness captures the degree to which parasites in the new environment are a subset of the original parasite community (Fig. 1, Patterson, 1987) and turnover measures the addition of new parasite species (Fig. 1). We predicted that parasite communities in captive primates would be a nested subset of the wild parasite community, with notable absences including native range parasites that require intermediate hosts or vectors. We also predicted that primates would acquire new parasites in captivity, especially parasites transmitted by close or non-close contact, for which transmission opportunities might exist in captive settings, as well as parasites known to infect humans. 

#### 2. MATERIALS AND METHODS

2.1 Data collection

We collected captive nonhuman primate parasite occurrence data from the primary literature using studies published between 1920 and 2012. We focused on 22 primate species representing 

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142	the four major primate lineages based on an initial list of primate species that were sampled well
143	for parasites in the wild, and that were known to be housed in captive settings (Table S1).
144	Primate species' scientific names, including synonyms, were based on well-accepted mammal
145	taxonomy (Wilson & Reeder, 2005). Data on parasites from captive primates were collected by
146	systematically searching the Web of Science (https://webofknowledge.com/), National
147	Agricultural Library (AGRICOLA, http://agricola.nal.usda.gov/), and PubMed
148	(http://www.ncbi.nlm.nih.gov/pubmed) databases. The search strategy involved use of general
149	parasite search terms with host scientific name (i.e., "species name" AND (parasit* OR
150	pathogen* OR disease OR infect* OR arthropod OR bacteria OR helminth OR fungi OR
151	protozoa OR virus OR vector)). Captive settings included zoological parks, wildlife
152	rehabilitation centers, animals kept as pets, and captive colonies used for behavioral or
153	biomedical research. We removed all cases of experimental infections and challenges, retaining
154	only reports of naturally occurring infections in captive settings, resulting in data from 241
155	sources. Comparable data on parasite infection from wild populations of the same set of primate
156	species were obtained from the Global Mammal Parasite Database (GMPD (Nunn & Altizer,
157	2005; Stephens et al., 2017). Data from the GMPD included 359 sources, are publicly available
158	(https://parasites.nunn-lab.org/), and have been used in numerous analyses of parasitism in wild
159	primates (e.g., Altizer, Nunn, & Lindenfors, 2007; Altizer et al., 2003; Cooper et al., 2012;
160	Dallas, Huang, Nunn, Park, & Drake, 2017; Davies & Pedersen, 2008; Nunn et al., 2004; Park et
161	al., 2018). For each host, parasites were only included if they were identified to the genus level at
162	least. To avoid potential double counting, parasites that were not identified to the species level
163	were omitted if a congener with a species epithet was present.

We recorded the transmission strategy of each parasite into five non-mutually exclusive categories (Pedersen, Altizer, Poss, Cunningham, & Nunn, 2005): close contact, non-close contact, vector-borne, sexually transmitted, and intermediate hosts. Parasites categorized as spread by close contact were communicable by close proximity or direct contact such as biting. scratching, mating contact, or other touching. Sexually transmitted parasites were a subset of close-contact transmitted parasites that are spread during copulation. Non-close contact involved transmission via fomites or contact with contaminated soil or water (which could include fecal-oral transfer). Vector-borne parasites were those spread via biting arthropods (ticks, mites, fleas, flies, and other invertebrates). Parasites transmitted by intermediate hosts have complex life cycles typically characterized by trophic transmission, and primates could serve as either intermediate or final hosts, or dead-end hosts. Parasites could exhibit more than one transmission mode (e.g., sexually transmitted parasites may also be transmitted by close-contact, and many parasites transmitted by close contact can also be transmitted by non-close contact). We also recorded whether the parasite species were known to infect humans and/or were zoonotic based on the known human parasites (Center for Disease Control, www.cdc.gov, Taylor, Latham, & Mark, 2001). 

2.2 Statistical Analyses

PSR estimates can be influenced by sampling effort, defined as the degree to which each host species or population has been studied for parasites (Altizer et al., 2003; Poulin, 1998). We accounted for variation in sampling effort between wild and captive hosts using three approaches. First, we compared the number of parasite studies available for each primate host species in the wild and under captive conditions. We used the smaller number of studies to randomly subsample the condition with the larger number of studies, and used rarefaction to 

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187 calculate the PSR expected if sampling efforts were equal between wild and captive conditions 188 (e.g., Colwell et al., 2012). To obtain standard errors on estimates of PSR and quantify intraspecific variation due to variation among studies, we bootstrapped the studies 1,000 times. 189 190 Second, we also rarefied both conditions to one less study than the total number of studies using 1,000 bootstrap replicates to obtain PSR and the standard deviation in PSR. We used the 191 192 specaccum function in the package vegan (Oksanen et al., 2013) in the R statistical environment (R Core Team, 2014) to conduct the rarefaction. In these analyses, Trachypithecus cristata was 193 omitted because this primate species had only one study of parasitism in captivity. As an 194 195 alternative, the third approach to correct for differences in sampling effort between conditions was to divide the observed PSR by the number of studies in our data set reporting on parasitism. 196 We employed a paired-sample *t*-test to investigate the hypothesis that PSR is higher in 197 198 wild versus captive hosts. To account for the statistical non-independence of species in comparative studies (Griffin & Nunn, 2011; Harvey & Pagel, 1991), including in paired 199 differences such as those used here (Lindenfors, Revell, & Nunn, 2010), we used a phylogenetic 200 paired *t*-test with the function *phyl.pairedttest* in the R package *phytools* (Lindenfors et al., 2010; 201 202 Revell, 2015). In addition to performing the paired *t*-test, this function provides an estimate of 203 phylogenetic signal,  $\lambda$ , which can range from 0 to 1. When  $\lambda = 0$ , this indicates that captive versus wild differences are unrelated to phylogeny, while  $\lambda = 1$  indicates that the difference 204

covaries with phylogeny as expected under a Brownian motion model of evolution (Freckleton,
Harvey, & Pagel, 2002; C.L. Nunn, 2011). We included the standard deviation or standard error
of the rarefied PSR in the *t*-tests. We downloaded a 50% majority rules consensus tree from the
posterior distribution of trees inferred using a supermatrix approach and a Bayesian inference
framework, available via the 10k trees project (Arnold, Matthews, & Nunn, 2010). We pruned

this tree to include only the 22 species in our study. We ran these t-tests using the three correctedestimates of PSR mentioned above.

To visualize parasite community similarity between captive and wild primates, we used principal components analysis (PCA) to summarize the matrix of parasite presence / absence in captive and wild hosts. We used the *prcomp* function in R to conduct a singular value decomposition of the original matrix, with each host having one row for captive and one row for wild presence / absence of each parasite species. Axes represent the maximum shared variance in parasite presence among hosts (Legendre & Gallagher, 2001). We retained the first two principal components based on the observation of decreasing variance explained by subsequent components in a scree plot. To determine if parasite transmission mode predicted separation in the two-dimensional space, we averaged the factor loadings for parasite species in each transmission mode. This approach is a way of visualizing the axes of variation in the species composition of communities (McGarigal, Cushman, & Stafford, 2000), and have been used to investigate diversity in microbial ecology (Dollhopf, Hashsham, & Tiedje, 2001) and in analyses of the microbiome (Clayton et al., 2016). 

Beta diversity was measured as the dissimilarity between wild and captive parasite communities, in which a value of 0 indicates that the communities shared exactly the same parasites and a value of 1 indicates that communities are completely different (i.e., sharing no species). The nestedness component of beta diversity reflects the loss of some species from the original wild community and the retention of others, while species turnover is due to new parasite species occurring in the captive population (Andrés Baselga, 2010). Values close to 1 for the turnover component reflect total change between parasite species found in the wild versus captivity, while values for the turnover component close to 0 indicate that all beta diversity is 

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due to nestedness. We computed the Sorenson index of beta diversity partitioned into the
 turnover component (Simpson index) and nestedness component, calculated in the R package
 *betapart* (Andres Baselga & Orme, 2012).

236 We tested the hypothesis that parasite species identity differed between wild and captivity 237 (dependent variables) depending on transmission mode (independent variable) using binomial 238 logistic regressions, with transmission mode coded as a binary (presence-absence) variable. Our 239 first set of models tested the hypothesis that parasites transmitted through vectors or intermediate hosts (independent variables) were not detected in captive animals (dependent variable). For this, 240 241 we divided the whole parasite dataset into separate subsets of parasites found in wild versus 242 captive primates. When a parasite present in a wild host species was absent from the 243 corresponding captive sample, that parasite was recorded as not found in captivity, otherwise it 244 was present. Our second set of models tested the hypotheses that parasites are more likely to be reported in captive environments (dependent variable) if they 1) exhibit close-contact 245 246 transmission, 2) exhibit non-close contact transmission, and 3) are zoonotic (independent 247 variables). Specifically, if a parasite species in the captive dataset was not present in the wild 248 dataset, we recorded that parasite to be newly detected in captivity. Logistic regressions were 249 conducted using the *glm* function in R, specifying a logit probability link. The significance of the 250 model was tested using the  $\chi^2$  statistic, implemented with the *anova* function in R. These 251 analyses were only run for the 13 primate species with sufficient numbers of parasite species 252 with variation in transmission mode.

To characterize how parasite traits predicted the occurrence of a parasite in captivity (whether carried over from the wild or newly acquired), we also calculated the proportion of parasites detected or not, across all host species and within each of the five transmission

categories (which as noted above are not mutually exclusive). We present these data per host
species and as means across species. We tested whether the proportion of parasites detected in
captivity differed by transmission mode using the phylogenetic paired-sample t-tests described
above.

260 Data accessibility: Data available from the FigShare Repository (to be archived upon261 acceptance of article).

## 

**263 3. RESULTS** 

The PSR of captive hosts was similar to that of their wild counterparts (phylogenetic mean difference in PSR = 1.55, n=22 host species, phylogenetic paired t-test: t = -0.82, p = 0.42,  $\sigma^2$  = 1.06, Table S1). Phylogenetic signal was low ( $\lambda = 0$ ). Species showed remarkable variation in whether the captive or wild host communities had higher PSR (Fig. 2). In one such example, Ateles paniscus had 31 parasite species reported from 12 captive studies, and 10 parasite species from 14 wild studies (Table S1). In general, the species accumulation curves for all species show that PSR rarely reaches an asymptote, but continues to rise with each additional study, indicating that there are many more host-parasite relationships to be discovered (Figs. S1-22). Results of t-tests were qualitatively similar across three analyses that used different corrections for differences in studies between conditions (Table S2). The principal component analysis to examine dissimilarity in parasite community composition resulted in two primary axes (Fig. 3, where each host species is represented by two points, one each for captive and wild settings, and points closer together have more similar parasite community composition than points that are farther apart). The first principal component 

axis (PC1) represented 22.55% of the variation in parasite community composition among hosts. The second principal component axis (PC2) represented 7.06% of the variation, and separated approximately half of the wild versus captive hosts: wild hosts largely exhibited positive values of PC2 and captive hosts tended to exhibit negative values (Fig. 3). The factor loadings represent how strongly each parasite species was correlated with each axis. When averaging the mean factor loadings among parasites according to their transmission mode, parasites with intermediate hosts loaded more strongly on the positive end of PC2 (-0.001) than parasites without intermediate hosts (mean = -0.11). 

Changes in parasite community composition between captive and wild host species pairs were predominantly due to the species turnover component of beta-diversity, with only a small contribution of the nestedness component for some host species (Fig. 4, Table S3). These results indicate that the species composition of parasite communities in the wild was nearly completely replaced with a different set of parasites in captivity. Two host species were exceptions to this pattern. The captive parasite community of silvered leaf monkeys (*Trachypithecus cristata*) was a nested subset of the wild community, and for orangutans (*Pongo pygmaeus*), nestedness made up approximately 30% of the beta diversity. 

Vector-borne parasites were significantly more likely to be found in the wild versus in captivity in six out of 13 primate hosts (Table 1). In only one host species were parasites with intermediate hosts significantly more likely to be reported from the wild than in captivity (Table 1). Parasites with close-contact transmission were significantly more likely to be detected in captivity in two out of 11 host species (Table 2). Non-close contact transmission was significantly more likely to be detected in captivity in one out of 11 hosts (Table 2). Parasites known to infect humans were not detected in captivity significantly more often than those that do

not infect humans (Table 2). We note, however, that a large percentage of parasites detected in both wild and captive primates are also known to infect humans (mean = 88%, range = 43 - 100%).

Across primate species, the mean percentage of vector-borne parasites reported from wild populations but not detected in captivity was 37.5%, compared to 28% with close-contact transmission, 36% with non-close contact transmission, and 4.5% with sexual transmission (Fig. 5). Surprisingly, the proportion of parasites transmitted via intermediate hosts that were known from the wild but were not detected in captivity was relatively low (mean = 13.7%), although the proportion of parasites with intermediate hosts present in the wild sample was also low (16% in the primate GMPD). On average, 60% of parasites only detected in captivity had close-contact transmission, 55.5% of parasites only detected in captivity had non-close transmission, and 15.5% were sexually transmitted (Fig. 5). Only 6.1% of parasites detected in captivity were vector-borne, and 14.0% of parasites detected in captivity had intermediate hosts. The proportion of parasites detected in captivity that had close-contact transmission was not significantly different from the proportion with non-close transmission, but was significantly higher for other transmission modes (Fig. 5).

#### 317 4. DISCUSSION

Primates are held in captivity for many purposes, ranging from biological research colonies to zoological parks and wildlife rehabilitation centers. Primates also have been relatively well sampled for parasites and pathogens, in part owing to their close relationships to humans, making them well-suited for analyses comparing parasites in wild and captive populations. Building on findings from invasion biology in which many native parasite species are lost when host species are introduced into new habitats (e.g., Torchin et al., 2003; Torchin & Mitchell, 2004), we 

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324 investigated predictions involving changes in parasite composition and richness in wild 325 compared to captive primates. Counter to our initial prediction that captive primates should 326 harbor fewer parasites than their wild counterparts, we found no significant difference in PSR 327 between captive and wild groups. Instead, our findings indicated that the number of parasites detected only in captive settings generally offsets those known only from the wild. In other 328 329 words, changes associated with captivity include the introduction of new parasites that replace 330 the loss of others. Despite similar richness estimates, the community composition of parasites 331 differed sharply between wild and captive primates. Rather than captive primate parasites being 332 nested subsets of those from wild primate hosts, the parasite communities of many wild hosts were almost completely replaced by a unique parasite community in captivity. 333 Differences in parasite community composition between wild and captive populations 334 335 can be partially explained by the dominant transmission mode of the parasite species. In 336 particular, parasites found exclusively in the wild were commonly transmitted by vectors such as 337 mosquitoes (Aedes sp.) and tsetse flies (Glossina sp.). We also predicted that parasite species 338 detected in captivity should be transmitted by close-contact or non-close transmission (e.g., 339 fecal-oral or contaminated substrates), but this prediction was only supported for two out of 13 340 host species. Across all 22 host species in this study, 60% of parasites not found in the wild but detected in captivity had close-contact transmission, while only 6% were vector-borne. It is 341 interesting that parasites transmitted by close and non-close contact were common in captivity, 342 343 despite regular medical care and hygiene practices. Collectively, these results illustrate that the

mode of parasite transmission is an important mechanism of parasite community change when

animals transition from wild to captive environments.

Table 3 highlights examples of key parasites we observed to be common in the wild (but not in the captivity) or common in captivity (but not in the wild) across parasite transmission modes. While parasite species presented in Table 3 do not infect *all* primates, they were common across many species in our dataset and are known to infect diverse hosts in captivity. The proportion of parasites that were identified exclusively in captivity and are known to infect humans ranged between 43 and 100% among primate species (Table 3, Taylor, Latham, & Mark, 2001). Notable pathogenic and zoonotic parasites in our captive primate dataset included protozoa such as Giardia duodenalis, nematodes such as Trichuris trichiura, bacteria such as *Mycobacterium tuberculosis*, and viruses such as Herpes simplex virus and human parainfluenza viruses (Table 3). These parasites are a known concern in captive populations, given the potential fatality of captive primates, the generality and host-breadth of the parasites, and their potential to spread to people (e.g., Stensvold et al., 2009). In future research, wild primates should be screened for parasites commonly found in captivity, because historically they have been under-appreciated in the wild (e.g., Blastocystis, Petrášová et al., 2011). Our results highlight important parasites to monitor in both captive and wild populations. PSR in wild primate species depends on host life history traits and ecological factors, such as geographic range area, social group size, foraging area, and population density (Nunn, 

Altizer, Jones, & Sechrest, 2003; Nunn et al., 2004). In most cases, it is not possible to directly compare the effects of these variables between wild and captive animals, either because the data were not consistently provided by the authors of the original paper (e.g., group size and population density) or the variables are simply not comparable in wild versus captive settings (e.g., geographic range size). If group sizes or cumulative habitat sizes of captive and wild populations differ, this might contribute to differences in parasitism (e.g., Guégan, Morand, &

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Poulin, 2005; Poulin, 2014). If group sizes were artificially larger or smaller in captivity than in the wild, this could cause parasite species richness in captivity to deviate from wild conditions. However, because some factors such as group size likely have stronger effects on infection prevalence than on parasite species richness, we do not believe this would bias our results (Rifkin, Nunn, & Garamszegi, 2012). In addition, the captive setting itself could lead to variation in parasitism; relevant variables include whether housing was indoor vs. outdoor, the number and types of other animal species in the facility, and changes in husbandry practices over time. Again, these data were not consistently reported in the papers on captive primates, and would most likely add random noise, not systematic bias. In a comparative study such as ours, including over 550 parasites from 600 studies, differences in methodology among studies could impact our estimates of PSR. For example, no single study quantified the total PSR of a particular host; instead, studies typically focus on a group of parasites, such as helminths, gut or blood-borne protozoa, or viruses. Detection methods employed across studies have different sensitivities, and not all studies were able to identify parasites to the species level, or discern closely related species (e.g., Entamoeba histolytica vs. E. *dispar*). Infection statuses for viruses and bacteria are often inferred from serology or from molecular screening, whereas helminth and gastrointestinal protozoan infections are assessed from fecal examination following flotations or fecal smears. We do not expect that including results from serology and molecular techniques will adversely affect the results, because similar

global analyses of parasite infection did not detect differences in results when omitting serologybased data (Olival et al., 2017; Pandit et al., 2018). Although these factors may limit the depth of interpretation of the results, we have no reason to expect that they will cause systematic biases across hosts and parasites that favor any particular hypotheses that we tested. Instead, these

confounds should add random noise to the data, making patterns more difficult to detect, rather
than mislead us to accept a false positive result. Looking forward, several new approaches offer
opportunities for consistent sampling across species, which would revolutionize attempts to
understand broad patterns of parasitism. In particular, DNA barcoding and metagenomics
provide methods to consistently identify current infection (Besansky, Severson, & Ferdig, 2003;
Pallen, 2014). Given rapid advances in molecular techniques, a standardized procedure for
molecular identification of parasites may not be far off.

By providing a comparative context for understanding parasitism in wild and captive primates, our study reinforces the need for vigilance during reintroduction programs. Captive primates reintroduced to the wild could bring with them a number of parasites that are unique to the captive environment, with detrimental effects on the wild population (Viggers, Lindenmayer, & Spratt, 1993). Any animal intended for reintroduction should be guarantined and screened for disease agents before release, and individuals harboring pathogens should be cleared of infection or removed from reintroduction programs (e.g., callitrichid hepatitis in captive *Leontopithecus* populations, Viggers et al., 1993). Similarly, reintroduced individuals that had never encountered parasites from wild environments could be highly susceptible to naturally-occurring infections. Both of these patterns should be assessed when planning translocation programs (Jana Petrášová et al., 2010). For example, moose and caribou reintroductions in North America suffered owing to the spread of a meningeal worm from sympatric white-tailed deer (Anderson, 1972), and reintroduced whooping cranes exhibited high mortality due to eastern equine encephalitis virus spread by mosquito vectors (Carpenter, Clark, & Watts, 1989). In an effort to restore wild populations of golden lion tamarins (*Leontopithecus rosalia*), individuals raised in captivity were released to the wild, and half died of an unknown disease (Kleiman et al., 1986). It is crucial to 

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3 4	415	assess these factors in programs to reintroduce animals from captivity to the wild (Baker, 2002;
5 6	416	Hartley & Sainsbury, 2017). Findings in this study highlight the specific parasites that are
7 8 9	417	common for each species and should be monitored.
10 11 12	418	In conclusion, parasite communities varied considerably between wild and captive
13 14	419	settings for 22 primate species, but without significant differences in the total number of parasite
15 16	420	species harbored by each group. Dissimilarity between wild and captive parasite communities
17 18	421	was driven more by parasite replacement rather than by net parasite loss. Replacement of vector-
19 20 21	422	borne parasites from the wild, and the addition of new close-contact and non-close transmitted
22 23	423	parasites in captivity, are potential threats to captive primates and present risk of spillover or
24 25	424	spill-back to humans. Our results also contribute to understanding the ecological drivers of
26 27 28	425	parasite communities, with applications for captive and wild management of primate disease
28 29 30 31	426	agents.
32 33	427	Acknowledgements
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36 37	429	entry during this project. We thank members of the Nunn lab for assistance with analyses,
38 39	430	especially R. Griffin, and helpful revisions on earlier versions of the manuscript. We also thank
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45 46	433	SA, JR, and DC designed the study and collected data. JPH and DC analyzed data, and all
47 48	434	authors contributed to writing and revising the manuscript.
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**Table 1.** Results of logistic regressions for each host species, predicting parasite species found in the wild and not reported in captivity, relative to parasite transmission mode. The number of parasite species lost in captivity is given out of the total wild parasite community. Rows in **bold** were significant at alpha = 0.05, tested against the  $\chi^2$  statistic.

		coefficient Intermediate-			
Host species	<b>Proportion of</b>	host		coefficient Vector	
binomial name	parasites	transmission	р	transmission	р
Aotus trivirgatus	10/13	17.59	0.28	-1.54	0.26
Chlorocebus					
aethiops	21/31	0.75	0.51	18.3	0.006
Colobus guereza	15/23	17.08	0.18	17.08	0.18
Erythrocebus patas	8/12	17.36	0.33	17.36	0.33
Gorilla gorilla	17/34	16.1	0.16	17.27	0.03
Macaca					
fascicularis	25/32	16.42	0.21	1.55	0.13
Macaca mulatta	27/33	-1.65	0.29	17.31	0.10
Mandrillus sphinx	15/23	1.25	0.26	18.34	0.03
Pan troglodytes	55/80	0.27	0.71	2.43	0.002
Papio					
cynocephalus	53/64	17.25	0.02	-0.76	0.43
Pongo pygmaeus	11/27	17.04	0.17	0.54	0.53
Saimiri sciureus	34/42	-0.55	0.57	19.16	0.0004
Trachypithecus					
cristata	19/24	17.4	0.22	19.1	0.007

**Table 2.** Results of logistic regressions for each host species, predicting parasite species reported in captivity but not in the wild, based 443 on parasite transmission mode. Rows in **bold** were significant at alpha = 0.05, tested against the  $\chi^2$  statistic.

Host species binomial name	Proportion of parasites	coefficient Zoonotic	р	coefficient Close contact	р	coefficient Environmental	р
Aotus trivirgatus	7/10	0.41	0.78	0.41	0.78	-18.82	0.10
Chlorocebus aethiops	18/27	1.39	0.18	-0.8	0.38	< 0.001	1.00
Erythrocebus patas	14/16	NA	NA	-17.96	0.26	-16.86	0.34
Gorilla gorilla	18/28	1.45	0.27	0.69	0.53	-0.62	0.45
Macaca fascicularis	51/59	1.17	0.16	-0.23	0.76	0.23	0.76
Macaca mulatta	53/59	0.27	0.82	-19.18	0.003	2.14	0.03
Mandrillus sphinx	7/16	NA	NA	-0.41	0.7	1.57	0.18
Pan troglodytes	26/49	0.59	0.54	-0.47	0.42	-0.67	0.26
Papio cynocephalus	19/30	1.16	0.25	0.59	0.44	-0.18	0.81
Pongo pygmaeus	28/44	0.62	0.56	0.41	0.54	1.1	0.09
Saimiri sciureus	18/26	18.66	0.99	2.05	0.02	-1.2	0.17

**Table 3.** Common primate parasites reported in the wild but not reported in captivity, or reported in captivity but not in the wild, characterized by their transmission modes. Parasites with an asterisk are known to infect humans (Taylor, Latham, & Mark, 2001). Note this is not an exhaustive list of all parasites in the dataset.

Transmission mode	Common in wild but not in captivity	Common in captivity but not in wild	
Close or non-close	Viruses	Bacteria	
contact	<i>Ebolavirus</i> sp.*	Streptococcus pneumoniae*	
	Simian immunodeficiency virus	Pseudomonas aeruginosa*	
	Bacteria	Mycobacterium bovis*	
	<i>Treponema</i> sp.*	Salmonella sp.*	
		Shigella flexerni*	
		Viruses	
		Deltaretrovirus STLV 2	
		Simian foamy virus	
		Human parainfluenza virus*	
		Herpes simplex virus*	
		Protozoa	
		Blastocystis hominis*	
		Cryptosporidium sp.*	
		Entamoeba histolytica*	
		Iodamoeba sp.*	
Vector-borne	Viruses	Bacteria	
	Yellow fever virus	Francisella tularensis*	
	Protozoa	Viruses	
	<i>Plasmodium</i> sp.*	Chikungunya virus*	
	<i>Trypanosoma</i> sp.*		

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3	Hepatocystis sp.
4	Helminth
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6	Loa loa*
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**Figure 1**. Schematic demonstrating the differences between nestedness and turnover components of beta diversity. Nestedness results when parasites in wild hosts are not present in captivity. Turnover results when parasite species are different between wild and captive hosts. Both nestedness and turnover can occur in a host to varying degrees, and make up the beta diversity between the wild and captive environments.

**Figure 2**. Plot of rarefied parasite species richness (PSR) in 21 paired wild and captive primate species. There was no significant difference in rarefied species richness between captive and wild conditions (phylogenetic paired-sample t-test). PSR was rarefied by the minimum number of studies in either the wild or captive host. The size of the circle is proportional to the number of studies (log-transformed). Species in blue had lower PSR in captivity than in the wild, while species in red had higher PSR in captivity than the wild. Data were offset slightly to allow visualization of overlapping points.

**Figure 3**. Principal components summarizing the host-parasite matrix in two dimensions. Every point in the plot is a captive or wild host and the distance among points illustrates their dissimilarity in parasite community composition. The second component, which discriminates captive and wild parasite communities, is characterized by parasites without intermediate hosts having negative factor loadings. NWM = New World monkeys, OWM = Old World monkeys.

**Figure 4**. Boxplots representing the two components of beta diversity, nestedness and turnover between parasite communities of wild and captive primates. A) The turnover component of beta diversity (Simpson's index), B) The nestedness component (SNE), C) The overall beta diversity (Sorenson's index).

**Figure 5.** Comparison of the proportion of parasite species known from the wild but not detected in captivity (green triangles) or not reported in the wild but detected in captivity (orange circles) by parasite transmission mode. Points represent the mean proportion, and bars represent 95% confidence intervals. Note that transmission modes are not mutually exclusive (e.g., sexually transmitted parasites exhibit close contact transmission, and some parasites exhibit both close and non-close contact transmission). The proportion of parasites detected in captivity that had close-contact transmission was significantly higher than for parasites with vector-borne transmission (t = -3.24, p = 0.005,  $\lambda = 0.52$ ), intermediate-host transmission (t = -3.22, p = 0.005,  $\lambda = 0.22$ ), and sexual transmission (t = -9.98, p < 0.001,  $\lambda = 0$ ), but was not significantly higher when compared to non-close transmission (t = 0.43, p = 0.67,  $\lambda = 0.41$ ).

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### **Research highlights:**

- Changes in host environments—from wild to captive—can lead to changes in parasitism; studying these changes can inform captive wildlife management, wildlife relocation programs, and zoonotic disease risk assessment.
- Comparing 22 species of primates, we found high parasite species turnover in captive hosts, but no overall difference in parasite richness between wild and captive populations. Captive primates had fewer vector-borne parasites, and appeared to gain parasites transmitted via environmental exposure and close contact.
- Parasitism is an important consideration in translocating primates from the wild, and in reintroduction programs, owing to the potential for novel parasite transfers into human and wild primate populations.

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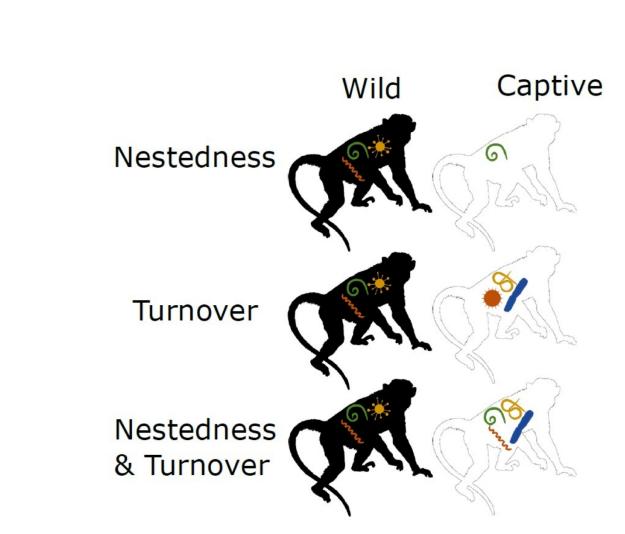


Figure 1. Schematic demonstrating the differences between nestedness and turnover components of beta diversity. Nestedness results when parasites in wild hosts are not present in captivity. Turnover results when parasite species are different between wild and captive hosts. Both nestedness and turnover can occur in a host to varying degrees, and make up the beta diversity between the wild and captive environments.

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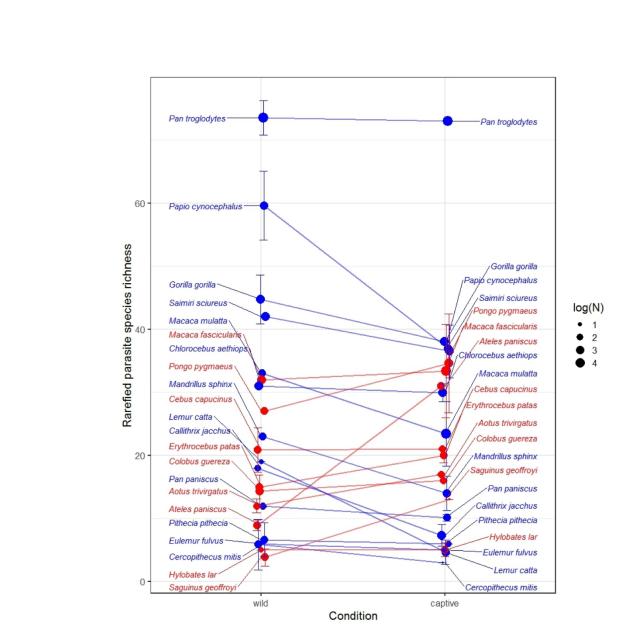
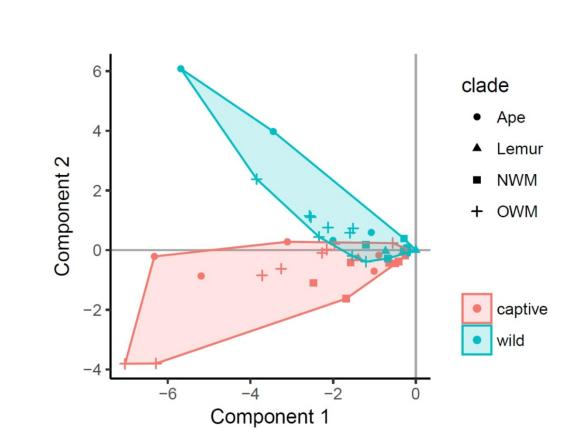
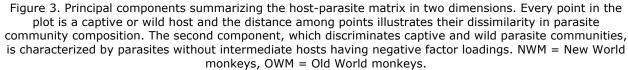


Figure 2. Plot of rarefied parasite species richness (PSR) in 21 paired wild and captive primate species. There was no significant difference in rarefied species richness between captive and wild conditions (phylogenetic paired-sample t-test). PSR was rarefied by the minimum number of studies in either the wild or captive host. The size of the circle is proportional to the number of studies (log-transformed). Species in blue had lower PSR in captivity than in the wild, while species in red had higher PSR in captivity than the wild. Data were offset slightly to allow visualization of overlapping points.

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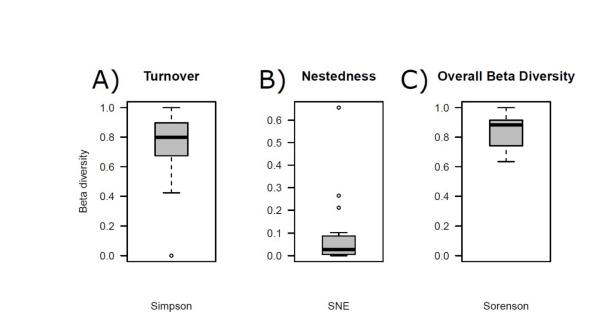


Figure 4. Boxplots representing the two components of beta diversity, nestedness and turnover between parasite communities of wild and captive primates. A) The turnover component of beta diversity (Simpson's index), B) The nestedness component (SNE), C) The overall beta diversity (Sorenson's index).

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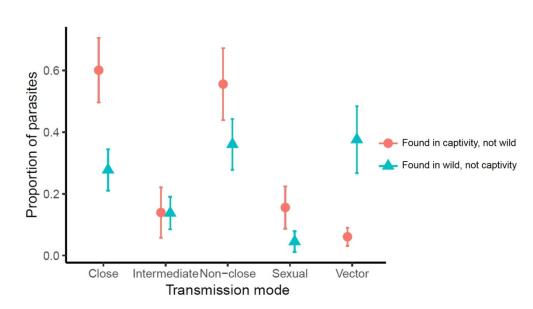


Figure 5. Comparison of the proportion of parasite species known from the wild but not detected in captivity (green triangles) or not reported in the wild but detected in captivity (orange circles) by parasite transmission mode. Points represent the mean proportion, and bars represent 95% confidence intervals. Note that transmission modes are not mutually exclusive (e.g., sexually transmitted parasites exhibit close contact transmission, and some parasites exhibit both close and non-close contact transmission). The proportion of parasites detected in captivity that had close-contact transmission was significantly higher than for parasites with vector-borne transmission (t = -3.24, p = 0.005,  $\lambda$  = 0.52), intermediate-host transmission (t = -3.22, p = 0.005,  $\lambda$  = 0.22), and sexual transmission (t = -9.98, p < 0.001,  $\lambda$  = 0), but was not significantly higher when compared to non-close transmission (t = 0.43, p = 0.67,  $\lambda$  = 0.41).

324x175mm (300 x 300 DPI)

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2	host	•		n.studies.captive	
3	Cercopithecus_mitis	3	26	2	
4	Chlorocebus_aethiops	33	31		30
5	Erythrocebus_patas	21	15	13	12
6 7	Eulemur_fulvus	5	6	3	3
8	Lemur_catta	23	19	21	3
9	Ateles_paniscus	31	10	12	14
10	Aotus_trivirgatus	17	13	9	10
11	Callithrix_jacchus	28	18	25	6
12	Cebus_capucinus	21	24	9	11
13	Saguinus_geoffroyi	13	18	2	11
14	Saimiri_sciureus	42	42	32	
15 16	_ Gorilla_gorilla	38	54	27	35
10	Hylobates_lar	14	5	13	4
18	Pan_paniscus	11	12		10
19	Pan_troglodytes	73	80	67	76
20	Pongo_pygmaeus	58	27	28	15
21	Pithecia_pithecia	6	13	5	12
22	Macaca_fascicularis	79	32	68	24
23	Macaca_mulatta	91		74	
24 25	Mandrillus sphinx	21	23	17	11
26	Papio_cynocephalus	37	64	20	22
27	Colobus_guereza	16	22	0	17
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1			Development				
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3	1.5	2.1666667		0	2	24.495	
4	0.9428571	1.0333333		0.59	34	30.099	
5	1.6153846	1.25	19.97	1.09	12	13.92	1.82
6 7	1.6666667	2	4.2	1.788	2	4	1.87
7 8	1.0952381	6.3333333	22.28	1.18	20	13.6	4.93
9	2.5833333	0.7142857	28.51	3.29	11	9.47	0.63
10	1.8888889	1.3	15.09	1.75	8	11.98	1.11
11	1.12	3	26.98	0.74	24	15.5	5.16
12	2.3333333	2.1818182	18.69	1.69	8	22.55	2.46
13	6.5	1.6363636		0	2	16.513	
14	1.3125	1.5555556		2.08	31	40.798	
15	1.4074074	1.5428571		1.23	26	52.97	
16 17	1.0769231	1.25		0.77	12	4.26	
17	1.0705251	1.2		0.53	10	11.029	
19	1.0895522	1.0526316		1.045	66	79.259	
20	2.0714286	1.0520510		2.96	27	25.84	
21	2.0714280	1.0833333		0.403	4	12.33	
22		1.3333333					
23	1.1617647			1.91	67	31.12	
24	1.2297297	2.3571429		1.2	73	30.92	
25	1.2352941	2.0909091		1.44	16	21.21	
26	1.85	2.9090909		2.46	19	61.745	
27	1.777778	1.3529412	13.99	2.29	7	21.82	1.08
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1 2	Rare1wildN	Rare2captif	Rare2captive:	Rare2captive	Rare2wild	Rare2wildSD	Rare2wildN
3	11	•	. 0	2	5.87		
4	29	29.94	1.42	30	31		30
5	11		1.09	12	15	0	12
6	2		0	3	6	0	3
7	2		1.94	3	19		
8	13		0	12	8.92		12
9 10	9		0	9	11.98		9
10	5		1.71	6	11.50		
12	10		0	9	20.87		9
13	10		0	2	3.839		2
14	26		4.19	27	5.839 42		
15			4.19	27			
16	34				44.77	3.88	27
17	3		1.12	4	5	0	4
18	9		0.533	10	12		10
19 20	75	73	0	67	73.507		67
20 21	14		7.842	15	27		
22	11		0	5	6.623		5
23	23		7.4	24	32		24
24	13	23.45	5.12	14	33	0	14
25	10	14.014	2.7179	11	23	0	11
26	21	37	0	20	59.631	5.46	20
27	14	16	0	8	14.32	2.54	8
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